NET ECOSYSTEM PRODUCTION: A COMPREHENSIVE MEASURE OF NET CARBON ACCUMULATION BY ECOSYSTEMS

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Abstract. The conceptual framework used by ecologists and biogeochemists must allow for accurate and clearly defined comparisons of carbon fluxes made with disparate techniques across a spectrum of temporal and spatial scales. Consistent with usage over the past four decades, we define "net ecosystem production" (NEP) as the net carbon accumulation by ecosystems. Past use of this term has been ambiguous, because it has been used conceptually as a measure of carbon accumulation by ecosystems, but it has often been calculated considering only the balance between gross primary production (GPP) and ecosystem respiration. This calculation ignores other carbon fluxes from ecosystems (e.g., leaching of dissolved carbon and losses associated with disturbance). To avoid conceptual ambiguities, we argue that NEP be defined, as in the past, as the net carbon accumulation by ecosystems and that it explicitly incorporate all the carbon fluxes from an ecosystem, including autotrophic respiration, heterotrophic respiration, losses associated with disturbance, dissolved and particulate carbon losses, volatile organic compound emissions, and lateral transfers among ecosystems. Net biome productivity (NBP), which has been proposed to account for carbon loss during episodic disturbance, is equivalent to NEP at regional or global scales. The multi-scale conceptual framework we describe provides continuity between flux measurements made at the scale of soil profiles and chambers, forest inventories, eddy covariance towers, aircraft, and inversions of remote atmospheric flask samples, allowing a direct comparison of NEP estimates made at all temporal and spatial scales.

Key words: atmospheric CO_2 ; biosphere–atmosphere fluxes; carbon accumulation by ecosystems, measuring; carbon balance; disturbance; net biome production; net ecosystem production; net primary production; scaling.

INTRODUCTION

Some of the biogeochemical processes that affect the carbon balance of terrestrial ecosystems include photosynthesis (Farquhar et al. 1980, Collatz et al. 1991), plant respiration (Ryan 1991), microbial respiration (Parton et al. 1993), leaching losses (Neff and Asner 2001), erosion (Stallard 1998), herbivory (McNaughton et al. 1989), fire (Crutzen and Andreae 1990), icesheet expansion and retreat (Harden et al. 1992, Schlesinger 1997), and rates of rock weathering (Berner 1993). Human appropriation and modification of the earth's surface over the last several centuries has altered many of these processes, with consequences for net ecosystem carbon fluxes, atmospheric mass balance, and inputs to oceans (Vitousek et al. 1986, Houghton 1996, DeFries et al. 1999). With attention focused by the Kyoto Protocol (Schulze et al. 2000) and subsequent international dialogue on carbon emissions from individual nations, there has been an in-

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creasing need to define carbon budgets at regional and continental scales (Steffen et al. 1998). Unfortunately, direct measurements of carbon fluxes at regional to global scales are difficult with current technology (Tans 1993) so estimates of regional- and continental-scale C fluxes require the integration of remote atmosphere and satellite observations with field measurements and experimental manipulations (Running et al. 1999).

Initial investigations of biosphere-atmosphere carbon exchange made the assumption that the net flux can be approximated as the balance between photosynthesis and respiration (Keeling 1961, Machta 1972, Pearman and Hyson 1980, Fung et al. 1983). This was sensible given that many diurnal and seasonal patterns of atmospheric CO₂ concentration can be explained by considering only these two processes (Denning et al. 1996, Heimann et al. 1998). However, analysis of interannual and decadal dynamics in atmospheric CO₂ driven by changes within the terrestrial biosphere requires consideration of additional processes including fire, dissolved organic carbon (DOC) and dissolved inorganic carbon (DIC) losses in rivers, erosion, and land-use changes such as agriculture and timber harvest (Canadell et al. 2000, Pacala et al. 2001).

Concept	Acronym/ symbol	Global flux	Definition
Gross primary production	GPP	$\sim 100 - 150$ Pg C/yr	carbon uptake by plants during pho- tosynthesis
Autotrophic respiration	R _a	\sim ¹ / ₂ of GPP	respiratory loss (CO ₂) by plants for construction, maintenance, or ion uptake
Net primary production	NPP	$\sim \frac{1}{2}$ of GPP	$GP\dot{P} - R_{a}$
Heterotrophic respiration (on land)	$R_{ m h}$	\sim 82–95% of NPP	respiratory loss (CO ₂) by the hetero- trophic community (herbivores, mi- crobes, etc.)
Ecosystem respiration	R_{\circ}	~91–97% of GPP	$R_{\rm a} + R_{\rm b}$
Non-CO ₂ losses	c	~2.8–4.9 Pg C/yr	\dot{CO} , \dot{CH}_4 , isoprene, dissolved inor- ganic and organic carbon, erosion, etc; see Table 2.
Non-respiratory CO ₂ losses (fire)		~1.6-4.2 Pg C/yr	combustion flux of CO ₂ ; see Table 2.
Net ecosystem production	NEP	~±2.0 Pg Č/yr	total carbon accumulation within the ecosystem; defined in <i>Defining Net</i> <i>Ecosystem Production: A compre-</i> <i>hensive definition of NEP</i> .

TABLE 1. Summary of contemporary terrestrial ecosystem carbon fluxes.

In part because of the rapid expansion of carbon cycle analyses at regional to global scales, we are left with ambiguities in our conceptual framework (see Table 1 for a summary of ecosystem flux concepts). Should net ecosystem production (NEP), which was initially approximated in the first biosphere–atmosphere CO_2 studies as the difference between photosynthesis and respiration, formally include all carbon exchanges that influence net carbon accumulation by an ecosystem (as the name "*net* ecosystem production" implies)? Alternatively, should NEP refer solely to the photosynthesis and respiration components (as suggested by some recent analyses, e.g., Schulze and Heimann [1998], Steffen et al. [1998], and Buchmann and Schulze and [1999])?

DEFINING NET ECOSYSTEM PRODUCTION

Historical perspective

The simultaneous definition of NEP as the carbon accumulation within ecosystems and as the difference between gross primary production (GPP) and ecosystem respiration extends back in the ecological literature over four decades (Woodwell and Whittaker 1968, Woodwell and Botkin 1970, Reichle et al. 1975). These two definitions are equivalent when non-photosynthetic gains and non-respiratory losses to an ecosystem are negligible, and so in many applications over the last few decades the two definitions have been used interchangeably. Increasing interest in the global carbon budget and the partitioning of land and ocean carbon sinks in the late 1970s focused attention on the need to quantify non-respiratory losses from terrestrial ecosystems, including fire and river fluxes (Bolin et al. 1979). As described by Lugo and Brown (1986), even if the terrestrial biosphere were close to steady-state carbon balance, a substantial biosphere-atmosphere CO₂ sink would be required to match river carbon losses.

With new assessments of volatile organic compound (VOC), methane, fire, and river fluxes, the sum of non-CO₂ and non-respiratory losses from terrestrial ecosystems is substantial at stand, regional, and global scales. Combined, these fluxes represent a loss of $\sim 10\%$ of global net primary production (NPP) (Table 2), and thus provide motivation for defining NEP solely in terms of carbon accumulation at all scales of inquiry.

A comprehensive definition of NEP

To reconcile carbon flux measurements made with diverse techniques and across widely varying time and space scales, NEP must be defined as the rate at which carbon (C) accumulates within an ecosystem (i.e., the change in carbon storage over some time interval):

$$NEP = dC/dt.$$
 (1)

A critical element of this definition is that the ecosystem in consideration must have defined boundaries in three dimensions (it must be possible to enclose the ecosystem with a three-dimensional container or box). For example, in a forest ecosystem, the top of the box might be defined as the height of the tallest tree, while the bottom of the box might be defined as a specified soil depth. Another critical feature of this definition is that the start and end times of the measurement period (or interval of integration) must be specified. Conservation of mass requires that fluxes (F) across the ecosystem boundaries equal the rate of change in C within the ecosystem. Thus, NEP has the equivalent definition (Olson 1963):

$$NEP = F_{in} + F_{out}$$

$$\approx F_{GPP} + F_{R_e} + F_{fire} + F_{leaching} + F_{erosion}$$

$$+ F_{hydrocarbons} + F_{herbivory} + F_{harvest} + \cdots$$
(2)

Therefore, NEP encompasses all fluxes (F_i) across all the boundaries of the ecosystem, independent of the

Source	Components	Flux (Pg C/yr)	Flux range (Pg C/yr)	Reference
Non-CO ₂ losses				
Rivers	dissolved organic carbon	0.40	0.20-0.90	Schlesinger and Melack (1981), Degens (1982), Degens et al. (1991), Meybeck (1982), Suchet and Probst (1995), Stallard (1998)
	dissolved inorganic carbon	0.30	Ť	
	particulate organic carbon	0.30	Ť	
Volatile organic compounds, VOCs	isoprene monoterpene other reactive VOCs other non-reactive VOCs	0.50 0.12 0.26 0.26	+ + + +	Guenther et al. (1995)
Methane	natural sources anthropogenic biosphere	0.16 0.27	0.11-0.21 0.20-0.35	Prather et al. (1996)
Carbon monoxide, CO	fires photochemical oxidation of organic matter thermal oxidation of organic	1.0 0.06 0.04	0.50-1.50 0.03-0.09 0.01-0.08	Bergamaschi et al. (2000) Schade and Crutzen (1999) Schade et al. (1999)
N	matter			
Non-respiratory CO_2 losses		2.0		
Fires		3.0	1.6-4.2	Crutzen and Andreae (1990)
Total sum of non-CO $_2$ and non-respiratory CO $_2$ losses‡		6.6	4.4–9.2	

TABLE 2. Global non-CO₂ and non-respiratory carbon losses from terrestrial ecosystems.

† Range of estimates not available

[‡] Total flux is ~11% of global NPP at 60 Pg C/yr, which is ~6 times larger than the net terrestrial carbon flux estimated by Prentice et al. (2001).

driving mechanism or the degree of biological regulation (Fig. 1). This definition must include all fluxes because, in many instances, it is difficult to distinguish between C fluxes that are regulated solely by abiotic or biotic processes. Clear examples of fluxes with direct biological regulation include gross primary production (GPP) and ecosystem respiration (R_e ; including both autotrophic and heterotrophic components) in which CO₂ diffuses directly through a cell membrane. Fluxes associated with fire (F_{fire}) and soil erosion ($F_{erosion}$) also are strongly regulated by biological and ecological processes, though not necessarily at a cellular level. The



FIG. 1. Regulation of net ecosystem production (NEP) by processes in terrestrial ecosystems varies with the time–space scales. At longer timescales and larger spatial scales, dissolved organic and inorganic river fluxes (DOC/DIC), fires, and erosion play critical roles in regulating ecosystem carbon balance. On shorter timescales, gross primary production (GPP) and ecosystem respiration (R_e) are dominant processes.

fire flux ($F_{\rm fire}$; which includes CO₂, CO, CH₄, VOCs, and particulates) needs to be included in this definition, so that NEP is equivalent to the ecosystem mass rate of change over any timescale.

The role of non-CO₂ fluxes

Not all ecosystem C fluxes are in the form of CO₂ gas. Leaching (F_{leaching}) and hydrocarbon emissions $(F_{hydrocarbons})$ contribute significantly to NEP in many ecosystems. For example, in boreal peatlands and arctic tundra, vertical and lateral fluxes of dissolved organic carbon (DOC), dissolved inorganic carbon (DIC), and methane are as much as 5% to 10% of the net ecosystem-atmosphere CO₂ flux (Waddington and Roulet 1997, King et al. 1998, Reeburgh et al. 1998). In tundra ecosystems, the inclusion of dissolved CO₂ and methane fluxes in calculations of NEP reduced carbon uptake rates by 20% relative to the ecosystem-atmosphere CO₂ flux (Kling et al. 1992). Isoprene, terpene, and other volatile organic compound (VOC) emissions also contribute to NEP (Lerdau 1991, Monson et al. 1991, Guenther et al. 1995).

While stream fluxes of DOC are generally within the range of 0–10 g C·m⁻²·yr⁻¹ and average ~ 5 g $C \cdot m^{-2} \cdot yr^{-1}$, there is large variability (2–52) g $C \cdot m^{-2} \cdot yr^{-1}$) across ecosystems (Hope et al. 1994). Rates of DOC flux are also substantially higher in soils than in streams, indicating the need to explicitly define a vertical (depth) ecosystem boundary in NEP calculations (Neff and Asner 2001). The impact of soluble carbon fluxes on carbon balance may be especially important in streams draining boreal forests and in areas with substantial wetland cover (e.g., Moore 1989, Hope et al. 1994, Waddington and Roulet 1997). Particulate C losses can also be substantial in some ecosystems, particularly following disturbance (Bormann et al. 1974, Stallard et al. 1998). Are these losses significant when integrated to the global scale? A 1 Pg C/yr net terrestrial flux evenly distributed over all biomes corresponds to approximately a 10 g $C \cdot m^{-2} \cdot yr^{-1}$ flux. Thus, when integrated, stream and river losses are comparable in magnitude to the net terrestrial flux (Prentice et al. 2001).

Carbon can also be removed or delivered to an ecosystem by lateral transfer of organic material, as mediated by herbivores ($F_{herbivory}$) (McNaughton et al. 1989) or harvest ($F_{harvest}$) (Harmon et al. 1990). In many cases herbivores are an internal component of an ecosystem, and will not cause a net transfer of C across the three-dimensional shape defined for the purpose of the NEP measurement. In other cases, herbivores may transfer C across the predefined spatial boundaries of the ecosystem, and thus contribute to NEP. Possible situations where this might occur include mass migrations, insect outbreaks, or agricultural grazing, or in any situation where the defined ecosystem is small as compared to the size of a typical herbivore. These components have not been explicitly incorporated into previous definitions of NEP although they have been frequently inferred as a part of heterotrophic respiration (Parton et al. 1993). Either we need still another concept that reflects total ecosystem C accumulation or (as we argue here) we need to explicitly include all C exchanges in NEP, so that this term truly reflects the net C accumulation by ecosystems, as has been implied in the past.

Disturbance and NEP

All of the flux components that contribute to NEP defined in Eq. 2 are either a direct result of disturbance events or are responding to a complex history of multiple disturbances that leave a long-term biological legacy (Reichle et al. 1975).

"Disturbance" is frequently defined as a relatively discrete event that induces widespread mortality of the dominant species within an ecosystem (e.g., Aber and Melillo 1991). Climate variability, N deposition, and stimulation of plant growth by elevated levels of atmospheric CO_2 are often considered as separate processes (in terms consequences for C fluxes), though they also have the potential to cause stand-leveling mortality events (Neilson 1993) and thus also have the potential to serve as agents of disturbance.

As with the definition of NEP described above, a comprehensive definition of disturbance should be based on clearly defined spatial boundaries and time intervals. Here we suggest following the framework developed by Pickett et al. (1989). Pickett et al. (1989) define the ecological concept of disturbance as a change in the minimal structure of a system caused by an external agent. Applying these concepts at the ecosystem scale, minimal structure includes the species composition, the distribution of plant functional types, and canopy architecture. Stand-leveling wind storms, fires, mortality induced by insect outbreaks, stand-killing droughts, and harvesting by humans all fundamentally alter this minimal structure and originate from a different level of ecological organization, thus qualifying as external agents. At the biome scale, a change in the disturbance regime from climate change or humans may constitute an external agent, with the ensemble of vegetation types and stand ages defining the minimum structure.

Human impacts on NEP

Humans affect all of the NEP components on the right-hand side of Eq. 2 by modifying the atmospheric composition, nutrient levels, climate, erosion rates, and disturbance regime. Examples of fluxes associated with humans include the removal of crops in agricultural ecosystems for use in distant feedlots and cities, the lateral transfer of wood from forests to paper mills, from paper mills to suburban and urban areas, and then to landfills, and the altered heterotrophic respiration of plowed soils when disturbance reduces the protection

of soil organic carbon from microbial attack (Kurz et al. 1995, Barlaz 1998, Skog and Nicholson 1998).

The dominant components of NEP vary with scale

As indicated in the discussion above, NEP potentially involves different processes and forms of carbon (i.e., not just CO_2). These processes are best represented and studied at certain scales. For example, disturbances such as fire kill individual plants and consume individual detritus parts, but are best considered in the NEP context at the level of landscapes or at the stand scale over very long time periods relative to the disturbance return interval. This is in part because the probability of a fire occurring increases with the extent of the temporal or spatial scale. At finer scales fires are irregular and they appear to add an unreasonable and often misleading variance in NEP. We suggest that this feature of NEP could be addressed by explicitly defining the level that NEP is being reported by use of subscripts. This would then clarify the processes that are usually included and excluded from the analysis. Any scaling of NEP from one level to another would then involve the addition of the processes that are most appropriately studied at that scale. For example, NEP at the level of individual stands (NEP_{stand}, with length scales roughly from 1 m to 1 km) would require study of fluxes associated with GPP (F_{GPP}), ecosystem respiration (F_{Re}), hydrocarbons ($F_{hydrocarbons}$), leaching ($F_{leaching}$), and erosion (F_{erosion}) , although over short time spans the latter three terms might be neglected for some purposes. Fluxes associated with fire (F_{fire}), herbivory ($F_{\text{herbivory}}$), and harvest $(F_{harvest})$ may not be explicitly addressed at this scale (unless a disturbance occurs during a sampling interval). They would be seen at this scale as rare events that export C from the system. At larger scales, however, some of these processes would be seen as internal transfers while others would emerge as substantial contributors to NEP (Fig. 1).

Regional-scale estimates of NEP (NEP_{reg}, with length scales roughly from 1 km to 10^2 km) would include fluxes from a mosaic of stands with different disturbance histories and intensities. At this level, NEP is strongly regulated by direct CO₂ losses from fire and other disturbances as well as fluxes associated with GPP and R_e at various times following disturbance. It would include CO₂ emissions from crop burning, CO₂ and organics that enter groundwater and are subsequently emitted from lakes and streams, and CO₂ emitted from landfills and feedlots that was derived from NPP transported laterally from other ecosystems.

For the case of biome-level NEP (NEP_{biome}, with length scales roughly from 10^2 km to 10^4 km), fires, deforestation, erosion, and river DOC are significant contributing processes, even on relatively short time scales (months to years). At this scale, the impact of any individual disturbance event may be small compared with the ensemble of disturbance events that are simultaneously occurring, and human- and climate-in-

duced changes in the disturbance regime become critical regulators of NEP (Schulze and Heimann 1998, Canadell et al. 2000).

At the global scale, the integral of NEP over the entire land surface (NEP_{global}) represents the change in the total mass balance of the terrestrial biosphere, including plants, soils, herbivores, etc. NEP at this scale represents a transfer of C to ocean, atmosphere, and lithosphere reservoirs. By definition all transfers are internal, except the net flow to the atmosphere, oceans, or lithosphere. At the global scale, a clear definition of the borders of the terrestrial system is difficult given the dynamic mixing processes and CO_2 fluxes that occur at coastal margins and in estuaries.

The minimal scale for NEP

Use of a multiscale definition of NEP raises the issue of whether there is a minimal scale that emerges from NEP as a property. We suggest that this minimum scale is the same as the minimum scale for the definition of an ecosystem (Odum 1959). While most ecosystems include some combination of heterotrophs and autotrophs, the temporary or permanent absence of lightharvesting (photosynthetic) organisms clearly does not preclude application of the ecosystem concept neither should it preclude characterization of the ecosystem carbon balance with NEP.

MEASUREMENTS OF NEP AND BIOSPHERE-Atmosphere CO₂ Exchange

On short timescales (typically less than a decade), it is difficult to accurately measure changes in total ecosystem carbon storage against large and heterogeneous stocks of C in soils and vegetation using Eq. 1. More precision can be obtained by measuring fluxes across the boundaries of the ecosystem via Eq. 2 (Fig. 2). Use of Eq. 2 to estimate NEP requires that we identify and measure the dominant components of the flux. In practice, it also requires the conscious decision to neglect certain components, if they are believed to be small at the temporal or spatial scale of measurement.

At the time and space scale of eddy covariance flux measurements, the dominant one-way components of NEP are GPP and R_{e} . This has led to the definition of net ecosystem exchange (NEE) as the balance between GPP and R_e at half-hour to decadal time intervals (Wofsy et al. 1993). Defined in this way, NEE is a partial flux out the top of a three-dimensional box enclosing the ecosystem and is equivalent to NEP (Eq. 1 and 2) only at sites where other lateral and vertical C fluxes are demonstrably small. NEP at eddy flux tower sites in managed forests also includes any lateral removal of coarse woody debris in the footprint area, in addition to the ecosystem–atmosphere CO_2 flux. A tower in an eroding field might record a net uptake of C, even though NEP is at steady state, with ecosystem-atmosphere gains balanced by erosion losses (Harden et al. 1999). Moreover, the ecosystem boundaries should be



FIG. 2. The temporal and spatial domain of different techniques used to measure components of net ecosystem production (NEP) in terrestrial ecosystems. Atmospheric-inversion methods have been applied to contemporary flask CO_2 measurements as well as ice-core records extending over the last 1000 years (Joos et al. 1999, Rayner et al. 1999). Eddy covariance flux measurements from the Harvard Forest (Petersham, Massachusetts, USA) extend over one decade (Wofsy et al.1993).

clearly defined in stand- and regional-scale flux studies by disturbance age, erosion setting, and other key variables that affect NEP as it is "scaled up" to regional or global estimates (e.g., Rapalee et al. 1998).

For similar reasons, biosphere-atmosphere CO₂ fluxes inferred from atmospheric model inversions of CO₂ flask or aircraft measurements (Rayner et al. 1999) are also only partially representative of terrestrial NEP. At the global scale, river runoff contains DOC (dissolved organic C) and DIC (dissolved inorganic C) of mixed terrestrial and aquatic origin on the order of 0.2-0.9 Pg C/yr for DOC and 0.3 Pg C/yr for DIC (Schlesinger and Melack 1981, Degens 1982, Meybeck 1982, Degens et al. 1991, Suchet and Probst 1995). As previously stated, if the terrestrial biosphere were at steady state, then this net hydrologic C transport would require a one-way atmosphere-biosphere flux of the same magnitude (Lugo and Brown 1986). Estimates of C accumulation within the terrestrial biosphere based on atmosphere-biosphere fluxes must account for these hydrologic losses, and also assess the spatial domain and timescale over which the land-ocean flux returns to the atmosphere in coastal margins and in the open ocean. Atmospheric inversions of CO₂ will also fail to properly assign sources and sinks of total C when oxidation of CH₄, CO, and some VOCs occur in atmospheric regions that are offset from their terrestrial sources.

On timescales greater than a decade, it is possible to measure changes in total ecosystem C storage (Fig. 2). Chronosequences of stands of different ages show decade-to-century scale increases in C stocks following clearing (Richter et al. 1999), glacial retreat (Crocker and Major 1955, Harden et al. 1992), floodplain development (Yarie et al. 1998), fire (Harden et al. 2000), and large C losses associated with disturbance (Kasischke et al. 1995, Cohen et al. 1996). The C accumulation over any long time period depends on GPP (gross primary production) and the respiration of newly fixed carbon, respiration and other loss pathways for carbon that was fixed prior to the last disturbance, and carbon losses associated with the disturbance event. For example, NEP measurement following disturbance includes variable proportions of recent and old soil C (Trumbore and Harden 1997, Goulden et al. 1998).

NEP AND NET BIOME PRODUCTION

The new concept proposed by Schulze and Heimann (1998)

To address the issue of disturbance (which was widely neglected in initial biosphere-atmosphere modeling analyses and field measurement programs), Schulze and Heimann (1998) proposed the concept of net biome production (NBP), defined as the regional net carbon accumulation after considering C losses from fire, harvest, and other episodic disturbances. Specifically, the NBP concept acknowledges that small but consistent rates of C accumulation over most of the land surface (i.e., what occurs in most forests and grasslands) must be balanced by relatively infrequent, but large-magnitude releases associated with episodic disturbance (e.g., Rapalee et al. 1998). The infrequent nature of these release events makes it difficult to design terrestrial sampling programs that provide a true measure of the regional- or continental-scale C flux (Schulze and Heimann 1998).

In the definition of NBP, Schulze and Heimann (1998) distinguish between directional forcing (such as

changing levels of atmospheric CO₂ or temperature) and disturbance by episodic forcing (such as fires and harvest). According to Schulze and Heimann (1998), directional forcing affects NEP (net ecosystem production) fluxes through changes in NPP and microbial and herbivore respiration. In contrast, episodic disturbance is said to only affect NBP because of the different time scale and impact on ecosystem processes (frequently decimating aboveground and belowground biomass stocks) and the different measurement approach required. NBP is considered as a downstream flux at the landscape, regional, or continental scale, after stand-level fluxes of GPP, R_a , R_b , and NEP have been estimated. Here we question whether episodic disturbances that are included in NBP can be partitioned conceptually or practically from processes that are included in NEP.

Ambiguities introduced by the NBP concept

The distinction between NEP and NBP has limitations that compromise communication among biogeochemists, atmospheric scientists, and ecologists working at different scales. The limitations stem from three sources: (1) The definition of NBP assumes episodic disturbance and directional forcing can be distinguished from one another, allowing an unambiguous partitioning of fluxes between NEP and NBP. (2) The definition of NBP suggests that other terrestrial C fluxes (GPP, NPP, R_{e} , and NEP; Table 1) can be estimated (either measured or modeled) separately from episodic disturbance (and thus NBP). (3) The name and definition of NBP implies that episodic disturbance emerges only as a regulator of C fluxes at continental or "biome" scales, as a downstream process from NEP. We address these issues in the following paragraphs.

In many instances, episodic disturbances and directional forcing cannot be easily separated, yet our framework (and carbon accounting systems) must be rigorous enough to include this reality. For example, temperature increases could be classified as a directional forcing (and thus fall under NEP in the definition proposed by Schulze and Heimann [1998]). Consider the case, however, of a severe temperature or drought event that kills some of the vegetation. How much of this mortality must occur before the flux is considered NBP instead of NEP? If fire is an annual occurrence, as in many grassland ecosystems, or consumes only a small fraction of the vegetation, as in many ground fires, is this an episodic disturbance? Similar issues arise with herbivory and wind damage. Low levels of insect herbivory and wind damage are common in most ecosystems, while outbreaks and hurricanes may be relatively infrequent. If an outbreak or windstorm does occur, at what level of severity does it constitute an episodic disturbance vs. a directional one? Low levels of herbivory are usually treated as a component of NPP, but if certain levels of herbivory are counted in NBP one needs to decide which form goes with which process.

In the case of harvest, ecosystems may experience severe disturbance of the soil and canopy understory during clearing (Black and Harden 1995) or harvest may be restricted to removal of berries or removal of dead wood for fuel (Hao and Liu 1994). Should the harvest flux (associated with NBP according to Schulze and Heimann [1998]) include the changes in soil respiration triggered by harvest removal or subsequent soil erosion?

The implications of attempting to separate episodic and directional forcing are severe. Previous estimates of global NEP, based on the assumption that episodic disturbances are associated only with the NBP flux, are as high as 10 Pg C/yr into the land surface (Steffen et al. 1998, Prentice et al. 2001). Yet, as shown in the examples above, it is unclear as to exactly what fluxes should be included when NEP is defined in this way. It is essential that we separate our theoretical paradigm of terrestrial C balance from our ability (or inability) to accurately measure the net flux and its components. From a practical perspective, NEP and NBP are impossible to measure separately (it is impossible to make a pure measurement of NEP following the NBP-NEP distinction presented in Schulze and Heimann [1998], Buchmann and Schulze [1999], and Schulze et al. [2000]).

Disturbance is an integral and defining element of all ecosystems. Therefore measurements of GPP, NPP, $R_{\rm a}$, and $R_{\rm b}$ are impossible to consider outside the context of both episodic disturbance and directional forcing. For example, canopy photosynthetic capacity depends strongly on leaf nitrogen, which in turn, depends on soil N availability and the cumulative history of disturbance events that have precipitated N loss (Field and Mooney 1986, Schulze et al. 1994). Likewise, ecosystem respiration fluxes critically depend on total ecosystem carbon stocks and their distribution among live, labile, chemically recalcitrant, and physically protected forms (Schimel et al. 1994). Again, past disturbance frequencies and intensities are critical regulators of the distribution and amounts of C in these forms. The definition of NBP does not emphasize the fundamental role of episodic disturbance in shaping "upstream" fluxes (GPP, NPP, R_a , and R_b); the primary effect of disturbance is assumed to occur at very large spatial scales (Buchmann and Schulze 1999: Fig. 1).

Within every square meter of an ecosystem, the net C balance (and all of the one-way components) is influenced by the cumulative history of multiple, previous disturbances. In the boreal forest, for example, logs remain after fire from trees that grew two or more fire cycles previously and are still decomposing. Harden et al. (2000) found that it was very difficult to distinguish between heterotrophic respiration from the decomposition of organic material exposed or killed during the last major disturbance and microbial respiration of decomposing leaf and root litter derived from living vegetation.

Reconciling NBP with NEP

Because of the ambiguities created by distinguishing between NBP and NEP, we suggest a conceptual framework where NBP is equivalent to a comprehensive definition of NEP (see *Defining NEP: A comprehensive definition,* above) at regional or global scales. Thus, NBP also represents the total mass balance of terrestrial C:

$$NBP = NEP = dC/dt.$$
 (3)

Is the concept of "net biome productivity" then necessary? Given the conceptual framework described by ecologists over the last few decades (Woodwell and Whittaker 1968, Reichle et al. 1975, Lugo and Brown 1986, Aber and Melillo 1991), NBP does not represent a quantity that is fundamentally different from NEP. However, the NBP concept is extremely useful because it highlights the role of rapid episodic fluxes in shaping NEP at very large scales and the challenges of extrapolating terrestrial C measurements made at individual sites (where these rapid episodic fluxes are not easily measured). NBP also highlights the non–negligible contribution of lateral (harvest) fluxes out of ecosystems that may be very difficult to quantify at individual sites.

While there is value in the NBP concept, we believe that it is impossible to attempt to partition fluxes from the terrestrial biosphere in terms of their origin as either arising solely from episodic disturbance or directional forcing from climate or other processes, or to make a unique distinction between NEP and NBP.

NEP, NBP, AND CARBON MANAGEMENT

The relevance of disturbance effects on carbon exchange is also important to the development of C emission restrictions in the Kyoto Protocol.⁷ Article 3.4 of the protocol includes the possibility that ecosystemmanagement activities focused on containing disturbances such as fire and pest outbreaks could be considered for carbon uptake credits. At the scale of countries or continents, disturbance-associated C fluxes are very large. For example, the direct loss of C during boreal forest fires is predicted to reach as high as 0.8 Pg C/yr over the next 30-100 yr (Kasischke et al. 1995). Reductions in the rates of these emissions could, in theory, qualify for C credits. The amount of C that could be affected by such policy decisions is significant. For example it has been suggested that a 5% reduction in fire-induced C losses in United States could yield a 0.5 Pg C/yr reduction in C emissions (Sohngen and Haynes 1997).

The critical issues involving NEP (net ecosystem production) and NBP (not biome production) estimates from a policy perspective are based on the need to document changes in C storage associated with management activities. For the reasons discussed in this

⁷ URL: (http://www.unfccc.de/resource/convkp.html)

manuscript, it is critical to clearly define the boundaries of the region considered for C balance and to evaluate the impacts of previous disturbances on current rates of C uptake or loss at all spatial and temporal scales. Without consideration of these two issues, the accounting requirements defined in the Kyoto Protocol for the inclusion of terrestrial C fluxes in a broader C restriction cannot be met.

CONCLUSIONS

1) A robust definition of net ecosystem production (NEP) should be based on a full ecosystem mass balance and include non-CO₂ and non-respiratory components of the C flux and clearly defined temporal and spatial boundaries. Defined in this way, NEP provides an unambiguous measure of change in C storage that is conceptually consistent across all spatial and temporal scales, from an individual plot to the entire terrestrial biosphere. In our view, definitions of NEP solely based on the difference between NPP and R_h or GPP and R_e encourages a perspective in which other C transfers are ignored at the local scale, and thus reconciling carbon mass balance with ocean and atmosphere reservoirs becomes difficult at the global scale.

2) Disturbance is an integral property of all ecosystems. It affects all one-way C fluxes including GPP, R_a , and R_h , and hydrologic fluxes at all spatial and temporal scales. It does not emerge as a regulator solely at regional or biome scales.

3) Net biome production (NBP) as previously defined by Schulze and Heimann (1998) cannot be distinguished from NEP in many instances because episodic disturbance is frequently impossible to distinguish from directional forcing and because most C fluxes respond to disturbance over a wide range of temporal scales.

4) Equivalent mass-balance definitions of the two terms "NBP" and "NEP" allows for a consistent treatment of carbon in political frameworks for taxation, and for efficient comparison of fluxes at various spatial and temporal scales and between models and field observations.

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LITERATURE CITED

- Aber, J. D., and J. M. Melillo. 1991. Terrestrial ecosystems. Saunders College Publications, Philadelphia, Pennsylvania, USA.
- Barlaz, M. A. 1998. Carbon storage during biodegradation

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of municipal solid waste components in laboratory-scale landfills. Global Biogeochemical Cycles **12**:373–380.

- Bergamaschi, P., R. Hein, M. Heimann, and P. J. Crutzen. 2000. Inverse modeling of the global CO cycle 1. Inversion of CO mixing ratios. Journal of Geophysical Research Atmospheres 105:1909–1927.
- Berner, R. A. 1993. Paleozoic atmospheric CO₂—importance of solar-radiation and plant evolution. Science 261:68–70.
- Black, T. A., and J. W. Harden. 1995. Effect of timber harvest on soil carbon storage at Blodgett Experimental Forest, California. Canadian Journal of Forest Research 25:1385– 1396.
- Bolin, B., E. T. Degens, P. Duvigneaud, and S. Kemp. 1979. The global biogeochemical carbon cycle. Pages 1–53 *in* B. Bolin, E. T. Degens, S. Kemp, and P. Ketner, editors. The global carbon cycle. John Wiley and Sons, Chichester, UK.
- Bormann, F. H., G. E. Likens, T. G. Siccama, R. S. Pierce, and J. S. Eaton. 1974. The export of nutrients and recovery of stable conditions following deforestation at Hubbard Brook. Ecological Monographs 44:255–277.
- Buchmann, N., and E.-D. Schulze. 1999. Net CO_2 and H_2O fluxes from terrestrial ecosystems. Global Biogeochemical Cycles **13**:751–760.
- Canadell, J. G., et al. 2000. Carbon metabolism of the terrestrial biosphere: a multitechnique approach for improved understanding. Ecosystems **3**:115–130.
- Cohen, W. B., M. E. Harmon, D. O. Wallin, and M. Fiorella. 1996. Two decades of carbon flux from forests of the Pacific Northwest. BioScience 46:836–844.
- Collatz, G. J., J. T. Ball, C. Grivet, and J. A. Berry. 1991. Physiological and environmental regulation of stomatal conductance, photosynthesis and transpiration: a model that includes a laminar boundary layer. Agricultural and Forest Meteorology 54:107–136.
- Crocker, R. L., and J. Major. 1955. Soil development in relation to vegetation and surface age at Glacier Bay, Alaska. Journal of Ecology **43**:427–448.
- Crutzen, P. J., and M. O. Andreae. 1990. Biomass burning in the tropics: impacts on atmospheric chemistry and biogeochemical cycles. Science 250:1669–1678.
- DeFries, R. S., C. B. Field, I. Fung, G. J. Collatz, and L. Bounoua. 1999. Combining satellite data and biogeochemical models to estimate global effects of human-induced land cover change on carbon emissions and primary productivity. Global Biogeochemical Cycles 13:803–815.
- Degens, E. T. 1982. Transport of carbon and minerals in major world rivers. Part 1. Proceedings of a Workshop Arranged by the Scientific Committee on Problems of the environment (SCOPE) and the United Nations Environment Programme (UNEP). Hamburg University, Hamburg, Germany.
- Degens, E. T., S. Kempe, and J. E. Richey. 1991. Biochemistry of major world rivers. SCOPE 42. John Wiley and Sons, Chichester, UK.
- Denning, S. A., D. A. Randall, G. J. Collatz, and P. J. Sellers. 1996. Simulations of terrestrial carbon metabolism and atmospheric CO₂ in a general-circulation model. 2. Simulated CO₂ concentrations. Tellus Serial B 48:543–567.
- Farquhar, G. D., M. J. R. Fasham, M. L. Goulden, M. Heimann, V. J. Jaramillo, H. S. Kheshgi, C. Le Quere, R. J. Scholes, and D. W. R. Wallace. 2001. The carbon cycle and atmospheric carbon dioxide. Pages 183–237 in J. T. Houghton, Y. Ding, D. J. Griggs, M. Noguer, P. J. van der Linden, X. Dai, K. Maskell, and C. A. Johnson, editors. Climate change 2001: the scientific basis. Contributions of Working Group 1 to the Third Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge, UK.
- Field, C. B., and H. A. Mooney. 1986. The photosynthesis– nitrogen relationship in wild plants. Pages 25–26 in T. J.

Givinish, editor. On the economy of plant form and function. Cambridge University Press, Cambridge, UK.

- Fung, I., K. Prentice, E. Matthews, J. Lerner, and G. Russell. 1983. Three-dimensional tracer model study of atmospheric CO₂: response to seasonal exchanges with the terrestrial biosphere. Journal of Geophysical Research Atmospheres 88:1281–1294.
- Goulden, M. L., S. C. Wofsy, J. W. Harden, S. E. Trumbore, P. M. Crill, S. T. Gower, T. Fries, B. C. Daube, S.-M. Fan, D. J. Sutton, A. Bazzaz, and J. W. Munger. 1998. Sensitivity of boreal forest carbon balance to warming. Science 279:214–217.
- Guenther, A., et al. 1995. A global-model of natural volatile organic-compound emissions. Journal of Geophysical Research Atmospheres 100:8873–8892.
- Hao, W. M., and M.-H. Liu. 1994. Spatial and temporal distribution of biomass burning. Global Biogeochemical Cycles 8:495–503.
- Harden, J. W., J. M. Sharpe, W. J. Parton, D. S. Ojima, T. L. Fries, T. G. Huntington, and S. M. Dabney. 1999. Dynamic replacement and loss of soil carbon on eroding cropland. Global Biogeochemical Cycles 13:885–901.
- Harden, J. W., E. T. Sundquist, R. F. Stallard, and R. K. Mark. 1992. Dynamics of soil carbon during deglaciation of the Laurentide ice sheet. Science 258:1921–1924.
- Harden, J. W., S. E. Trumbore, B. J. Stocks, A. Hirsch, S. T. Gower, K. P. O'Neill, and E. S. Kasischke. 2000. The role of fire in the boreal carbon budget. Global Change Biology 6(Supplement 1):174–184.
- Harmon, M. E., W. K. Ferrell, and J. F. Franklin. 1990. Effects on carbon storage of conversion of old-growth to young forests. Science 247:699–702.
- Heimann, M., et al. 1998. Evaluations of terrestrial carbon cycle models through simulations of the seasonal cycle of atmospheric CO₂: first results of a model intercomparison study. Global Biogeochemical Cycles **12**:1–24.
- Hope, D., M. F. Billett, and M. S. Cresser. 1994. A review of the export of carbon in river water: fluxes and processes. Environmental Pollution 84:301–324.
- Houghton, R. A. 1996. Terrestrial sources and sinks of carbon inferred from terrestrial data. Tellus Series B 48:420–432.
- Joos, F., R. Meyer, M. Bruno, and M. Leuenberger. 1999. The variability in the carbon sinks as reconstructed for the last 1000 years. Geophysical Research Letters 26:1437–1440.
- Kasischke, E. S., N. L. Christensen, and B. J. Stocks. 1995. Fire, global warming, and the carbon balance of boreal forests. Ecological Applications 5:437–451.
- Keeling, C. D. 1961. The concentration and isotopic abundance of carbon dioxide in rural and marine air. Geochimica et Cosmochimica Acta 24:277–298.
- King, J. Y., W. S. Reeburgh, and S. K. Regli. 1998. Methane emissions and transport by arctic sedges in Alaska: results of a vegetation removal experiment. Journal of Geophysical Research Atmospheres 103:29083–29092.
- Kling, G. W., G. W. Kipphut, and M. C. Miller. 1992. The flux of CO_2 and CH_4 from lakes and rivers in arctic Alaska. Hydrobiologia **240**:23–36.
- Kurz, W. A., M. J. Apps, S. J. Beukema, and T. Lekstrum. 1995. 20th century carbon budget of Canadian forests. Tellus Series B 47:170–177.
- Lerdau, M. T. 1991. Plant function and biogenic terpene emissions. Pages 121–133 *in* T. D. Sharkey, E. A. Holland, and H. A. Mooney, editors. Trace gas emissions by plants. Academic Press, San Diego, USA.
- Lugo, A. E., and S. Brown. 1986. Steady state terrestrial ecosystems and the global carbon cycle. Vegetatio **68**:83–90.
- Machta, L. 1972. Mauna Loa and global trends in air quality. Bulletin of the American Meteorological Society 53:402– 420.

- McNaughton, S. J., M. Oesterheld, D. A. Frank, and K. J. Williams. 1989. Ecosystem-level patterns of primary production and herbivory in terrestrial habitats. Nature 341: 142–144.
- Meybeck, J. M. 1982. Carbon, nitrogen, and phosphorus transport in world rivers. American Journal of Science **282**: 401–450.
- Monson, R. K., A. B. Guenther, and R. Fall. 1991. Physiological reality in relation to ecosystem- and global-level estimates of isoprene emissions. Pages 185–205 in T. D. Sharkey, E. A. Holland, and H. A. Mooney, editors. Trace gas emissions by plants. Academic Press, San Diego, USA.
- Moore, T. R. 1989. Dynamics of dissolved organic carbon in forested and disturbed catchments, Westland, New Zealand, 1. Maimai. Water Resources Research 25:1321–1330.
- Neff, J. C., and G. P. Asner. 2001. Dissolved organic carbon in terrestrial ecosystems: synthesis and a model. Ecosystems **4**:29–48.
- Neilson, R. P. 1993. Vegetation redistribution: a possible biospheric source of CO₂ during climate change. Water, Air, and Soil Pollution **70**:659–673.
- Odum, E. P. 1959. Fundamentals of ecology. W.B. Saunders Company, Philadelphia, Pennsylvania, USA.
- Olson, J. S. 1963. Energy storage and the balance of producers and decomposers in ecological systems. Ecology 44: 322–331.
- Pacala, S. W., et al. 2001. Consistent land- and atmospherebased US carbon sink estimates. Science 292:2316–2320.
- Parton, W. J., J. M. O. Scurlock, D. S. Ojima, T. G. Gilmanov, R. J. Scholes, D. S. Schimel, T. Kirchner, J.-C. Menaut, T. Seastedt, E. Garcia Moya, A. Kamnalrut, and J. I. Kinyamario. 1993. Observations and modeling of biomass and soil organic matter dynamics for the grassland biome worldwide. Global Biogeochemical Cycles 7:785–809.
- Pearman, G. I., and P. Hyson. 1980. Activities of the global biosphere as reflected in atmospheric CO₂ records. Journal of Geophysical Research **85**:4468–4474.
- Pickett, S. T. A., J. Kolassa, J. J. Armesto, and S. L. Collins. 1989. The ecological concept of disturbance and its expression at various hierarchical levels. Oikos 54:129–136.
- Prather, M., R. Derwent, D. Ehhalt, P. Fraser, E. Sanhueza, and X. Zhou. 1996. Radiative forcing of climate change. Chapter 2. Climate change 1995. The science of climate change. Contribution of Working Group 1 to the Second Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge, UK.
- Prentice, I. C., G. D. Farquhar, S. von Caemmerer, and J. A. Berry. 1980. A biochemical model of photosynthetic CO₂ assimilation in leaves of C₃ plants. Planta **149**:78–90.
- Rapalee, G., S. E. Trumbore, E. A. Davidson, J. W. Harden, and H. Veldhuis. 1998. Soil carbon stocks and their rates of accumulation and loss in a boreal forest landscape. Global Biogeochemical Cycles 12:687–701.
- Rayner, P. J., I. G. Enting, R. J. Francey, and R. Langenfelds. 1999. Reconstructing the recent carbon cycle from atmospheric CO_2 , $\delta^{13}C$, and O_2/N_2 observations. Tellus Series B **51**:213–232.
- Reeburgh, W. S., J. Y. King, S. K. Regli, G. W. Kling, N. A. Auerbach, and D. A. Walker. 1998. A CH₄ emission estimate for the Kuparuk River basin, Alaska. Journal of Geophysical Research 103:29005–29013.
- Reichle, D. E., R. V. O'Neill, and W. F. Harris. 1975. Principals of energy and material exchange in ecosystems. *In* W. H. van Dobben and R. H. Lowe-McConnell, editors. Unifying concepts in ecology. Dr. W. Junk B. V. Publishers, The Hague, The Netherlands.
- Richter, D. D., D. Markewitz, S. E. Trumbore, and C. G. Wells. 1999. Rapid accumulation and turnover of soil carbon in a reestablishing forest. Nature 400:56–58.

- Running, S., D. Baldocchi, D. Turner, S. Gower, P. Bakwin, and K. Hibbard. 1999. A global terrestrial monitoring network integrating tower fluxes, flask sampling, ecosystem modeling, and EOS satellite data. Remote Sensing of the Environment **70**:108–127.
- Ryan, M. G. 1991. Effects of climate change on plant respiration. Ecological Applications 1:157–167.
- Schade, G. W., and P. J. Crutzen. 1999. CO emissions from degrading plant matter. II. Estimate of a global source strength. Tellus Series B 51:909–918.
- Schade, G. W., R. M. Hofmann, and P. J. Crutzen. 1999. CO emissions from degrading plant matter. I. Measurements. Tellus Series B 51:889–908.
- Schimel, D. S., B. H. Braswell, E. A. Holland, R. Mckeown, D. S. Ojima, T. H. Painter, W. J. Parton, and A. R. Townsend. 1994. Climatic, edaphic, and biotic controls over storage and turnover of carbon in soils. Global Biogeochemical Cycles 8:279–293.
- Schlesinger, W. H. 1997. Biogeochemistry: an analysis of global change. Academic Press, San Diego, California, USA.
- Schlesinger, W. H., and J. M. Melack. 1981. Transport of organic carbon in the world's rivers. Tellus 33:172–181.
- Schulze, D. E., F. M. Kelliher, C. Korner, J. Lloyd, and R. Leuning. 1994. Relationships among maximum stomatal conductance, ecosystem surface conductance, carbon assimilation rate, and plant nitrogen nutrition. Annual Review of Ecology and Systematics 25:629–660.
- Schulze, E.-D., and M. Heimann. 1998. Carbon and water exchange of terrestrial ecosystems. Pages 145–161 in J. Galloway, and J. M. Melillo, editors. Asian change in the context of global change. Cambridge University Press, Cambridge, UK.
- Schulze, E.-D., C. Wirth, and M. Heimann. 2000. Managing forests after Kyoto. Science 289:2058–2059.
- Skog, K., and G. Nicholson. 1998. Carbon cycling through wood products: the role of wood and paper products in carbon sequestration. Forest Products Journal 48:75–83.
- Sohngen, B. L., and R. W. Haynes. 1997. The potential for increasing carbon storage in United States unreserved timberlands by reducing forest fire frequency: an economic and ecological analysis. Climatic Change 35:179–197.
- Stallard, R. F. 1998. Terrestrial sedimentation and the carbon cycle: coupling weathering and erosion to carbon burial. Global Biogeochemical Cycles 12:231–257.
- Steffen, W., I. Noble, J. Canadell, M. Apps, E.-D. Schulze, P. G. Jarvis, and the IGBP Terrestrial Carbon Working Group. 1998. The terrestrial carbon cycle: implications for the Kyoto Protocol. Science 280:1393–1394.
- Suchet, P. A., and J. L. Probst. 1995. A global-model for present-day atmospheric soil CO₂ consumption by chemical erosion of continental rocks (GEM-CO₂). Tellus Series B 47:273–280.
- Tans, P. P. 1993. Observational strategy for assessing the role of terrestrial ecosystems in the global carbon cycle: scaling down to regional levels. Pages 179–190 *in* C. Field, and J. Ehlringer, editors. Scaling physiological processes: leaf to globe. Academic Press, San Diego, California, USA.
- Trumbore, S. E., and J. W. Harden. 1997. Accumulation and turnover of carbon in organic and mineral soils of the BO-REAS northern study area. Journal of Geophysical Research Atmospheres 102:28817–28830.
- Vitousek, P. M., P. R. Ehrlich, A. E. Ehrlich, and P. A. Matson. 1986. Human appropriation of the products of photosynthesis. BioScience 36:368–373.
- Waddington, J. M., and N. T. Roulet. 1997. Groundwater flow and dissolved carbon movement in a boreal peatland. Journal of Hydrology **191**:122–138.
- Wofsy, S. C., M. L. Goulden, J. W. Munger, S. M. Fan, P. S. Bawkin, B. C. Daube, S. L. Bassow, and F. A. Bazzaz.

1993. Net exchange of CO_2 in a mid-latitude forest. Science **260**:1314-1316.

Woodwell, G. M., and D. B. Botkin. 1970. Metabolism of ecosystems by gas exchange techniques. Pages 73-85 in D. E. Reichle, editor. Analysis of temperate forest ecosystems. Springer-Verlag, New York, New York, USA. Woodwell, G. M., and R. H. Whittaker. 1968. Primary pro-

duction in terrestrial ecosystems. American Zoologist 8: 19-30.

Yarie, J., L. Viereck, K. Van Cleve, and P. Adams. 1998. Flooding and ecosystem dynamics along the Tanana riverapplying the state-factor approach to studies of ecosystem structure and function on the Tanana river floodplain. BioScience 48:690-695.